

## Soil Bacterial and Fungal Community Structure Across a Range of Unimproved and Semi-Improved Upland Grasslands

Nabla Kennedy, Suzanne Edwards and Nicholas Clipson

Microbial Ecology Group, Department of Industrial Microbiology, University College Dublin, Belfield, Dublin 4, Ireland

Received: 21 December 2004 / Accepted: 11 March 2005 / Online Publication: ■

### Abstract

Changes in soil microbial community structure due to improvement are often attributed to concurrent shifts in floristic community composition. The bacterial and fungal communities of unimproved and semi-improved (as determined by floristic classification) grassland soils were studied at five upland sites on similar geological substrata using both broad-scale (microbial activity and fungal biomass) and molecular [terminal restriction fragment length polymorphism (TRFLP), automated ribosomal intergenic spacer analysis (ARISA)] approaches. It was hypothesized that microbial community structure would be similar in soils from the same grassland type, and that grassland vegetation classifications could thus be used as predictors of microbial community structure. Microbial community measurements varied widely according to both site and grassland type, and trends in the effect of grassland improvement differed between sites. These results were consistent with those from similar studies, and indicated that floristic community composition was not a stable predictor of microbial community structure across sites. This may indicate a lack of correlation between grassland plant composition and soil microbial community structure, or that differences in soil chemistry between sites had larger impacts on soil microbial populations than plant-related effects.

### Introduction

Grassland plant communities can often be described in terms of the characteristic species that make up their composition. However, the extent of coupling between

grassland plant composition and soil microbial structure is poorly understood. It is thought that plants can influence microbial populations in soil through rhizodeposition, with changes in the rate and quality of substrate input occurring due to alterations in plant species composition and diversity [40, 50]. Plant species typical of upland grasslands have been shown to exhibit a selective effect on the microbial populations of their rhizospheres, with *Agrostis capillaris*, *Lolium perenne*, and *Trifolium repens* stimulating the growth of *Proteobacteria* [22, 41], whereas specific associations between dominant species of unimproved and improved upland grasslands and mycorrhizal fungi have been discovered [21, 44, 52, 51, 57, 64]. Shifts in overall microbial community structure have also been linked to plant species traits, with these changes usually attributed to differences in substrate deposition of different plants [3, 53, 61].

Exploration of links between plant and microbial community structure in grasslands is desirable because, although vegetation composition is relatively easy to qualify and quantify, soil microbial communities are much more difficult to characterize. Full characterization of soil fungal and bacterial populations requires multiple samplings and the use of expensive and time-consuming analyses, such as DNA-based community fingerprinting [28]. These approaches provide a detailed view of microbial community structure, but require much more time and effort than the vegetational surveys required to classify grasslands. If consistent links could be found between plant composition and soil microbial community structure, grassland vegetational classifications could be used as predictors of associated soil microbial structure.

In recent years, the National Vegetation Classification (NVC) system [49] has been successfully used to characterize plant communities in the UK and Ireland. Although broad classification systems have previously been used for grasslands [46], the NVC allows classifications into much more precise categories. Its applica-

Correspondence to: Nabla Kennedy at present address: Ecosystem Science and Management Programs, University of Northern British Columbia, 3333 University Way, Prince George, BC V2N 4Z9, Canada; E-mail: nablak@alum.colby.edu

tion to upland grasslands allows differentiation between unimproved grasslands (classified as U4a), and semi-improved grasslands (classified as U4b) [49]. Unimproved U4a grasslands are dominated by the plant species *Festuca ovina*, *A. capillaris*, and *Galium saxatile*. Semi-improved grasslands, although dominated by the same species as unimproved grasslands, also contain subcommunities of *L. perenne* and *T. repens*, indicating a limited degree of improvement [49].

This study attempts to determine if links between plant community composition and microbial community structure are consistent across a range of geographically separated sites with similar physical and chemical characteristics. The microbial community structure of unimproved (U4a) and semi-improved (U4b) grassland soils was studied at five upland sites on similar bedrock substrata in the Wicklow Mountains, Ireland, using both broad-scale (microbial activity and fungal biomass) and molecular [terminal restriction fragment length polymorphism (TRFLP), automated ribosomal intergenic spacer analysis (ARISA)] approaches. The central hypothesis was that microbial community structure would be similar in soils from the same grassland type, and that NVC grassland vegetation classifications could thus be used as predictors of microbial community structure.

## 104 Methods

105 **Field Sites.** Soil was taken from areas of unimproved (U4a) and semi-improved (U4b) grassland at field sites (Long Hill, Sally Gap, Lough Tay, Kings River, and Annagh Hill) in upland regions of counties Wicklow and Wexford in June 2000 (see Table 1 for grid references). The maximum longitudinal distance between sites was 45

km, and the maximum latitudinal distance between sites was 22 km. All sites were located on granite/quartzite bedrock substrata (as determined by reference to Geological Survey of Ireland map [20]), had similar peaty podzol soil composition (determined in consultation with soil survey bulletin map [17]), and contained areas of unimproved (U4a) and semi-improved (U4b) grassland. Floristic composition was analyzed according to the UK National Vegetation Classification (NVC) system [49]. Briefly, 1 m quadrats were placed randomly five times around the grassland area and plant species composition and estimates of abundance as percent cover were determined visually [49].

**Sampling Regime.** Three sampling points within each grassland type at each site were randomly selected, and soil samples were taken using a corer to remove three replicate cores (4 cm diameter, 10 cm depth) at each sampling point. As extensive grass roots were visible throughout all cores, all soil was assumed to have been in contact with plant roots and was considered rhizosphere. Soil was sieved to <4 mm to remove plant and root material, and stored at 4°C for less than 7 days for chemical analyses, microbial activity, and biomass analysis, and at -20°C for molecular analyses.

**Soil Physical and Chemical Analysis.** Soil samples were analyzed for pH<sub>water</sub> by an electrometric method [55], using a single junction reference electrode (Orion Instruments, Boston, MA, USA). Total nitrogen was determined by the Kjeldahl method [7], using a Kjeltac system 2000 Digestion apparatus and 2100 Distillation unit (Foss Tecator, Höganäs, Sweden). Soil was digested at 420°C for 1 h with a mixture of potassium sulfate, copper sulfate, and sulfuric acid, before undergoing steam

t1.1 **Table 1. Map grid references [Ordnance Survey of Ireland (OSI) grid coordinates] and chemical composition of soils from U4a and U4b grassland types at each field site (n = 3), and the average value for each soil type over all sites (n = 15)**

t1.2	OSI coordinates	pH		% Nitrogen		Phosphorus (mg P kg <sup>-1</sup> )		Potassium (mg K kg <sup>-1</sup> )		
		U4a	U4b	U4a	U4b	U4a	U4b	U4a	U4b	
t1.3										
t1.4	Site									
t1.5	Long Hill	O 218 124	3.82	4.57	0.97	0.72	12.8	13.7	241	366
t1.6	Sally Gap	O 116 124	4.23	4.12	0.44	0.47	8.2	12.4	42	43
t1.7	Lough Tay	O 160 087	3.46	3.95	0.68	0.71	11.5	11.6	167	241
t1.8	Kings River	O 008 014	4.36	3.92	0.61	0.72	11.1	12.2	157	173
t1.9	Annagh Hill	T 107 672	3.46	3.68	1.27	1.18	12.7	12.3	283	220
t1.10	Average		3.87	4.05	0.80	0.76	11.3	12.4	178	209
t1.11	SED (Site*Grassland type)		0.159		0.048		0.73		4.9	
t1.12	p Values									
t1.13	Site		***		***		***		***	
t1.14	Grassland type		*		NS		**		***	
t1.15	Site*Grassland type		***		***		**		***	

t1.16 Means and standard error of differences (SED) are shown. ANOVA p values for site, grassland type and their interaction are shown as NS, not significant; \*p < 0.05; \*\*p < 0.01; \*\*\*p < 0.001.

144 distillation. Nitrogen concentrations were determined after  
145 colorimetric titration with 0.1 M hydrochloric acid. Using  
146 ammonium sulfate as a standard, the recovery of nitrogen  
147 from these soil samples was estimated at  $82.8 \pm 2.6\%$ .

148 Extractable phosphorus levels in soil samples were  
149 determined using the Morgan extraction [45]. Phospho-  
150 rus was measured using the colorimetric ammonium  
151 molybdate–ascorbic acid method [36]. Absorbance was  
152 measured on a Jenway 6300 Spectrophotometer (Essex,  
153 UK) and compared to that of standard phosphorus  
154 solutions to determine extractable phosphorus concen-  
155 tration. Soil potassium concentrations were measured  
156 using the ammonium acetate centrifuge method [27].  
157 Phosphorus was extracted with ammonium acetate and  
158 analyzed by atomic emission spectrometry on a Corning  
159 410 flame photometer (High Wycombe, Bucks, UK),  
160 with sample readings compared to those of standard  
161 potassium solutions to determine potassium content.

162 *Total Microbial Activity.* Total microbial activity  
163 was assessed as triphenylformazan dehydrogenase activity  
164 and was determined based on a modification of the  
165 method of Thalmann [54], as previously described [8].

166 *Fungal Biomass.* Total soil ergosterol was quan-  
167 tified as described previously [9]. Briefly, 5 g soil was  
168 vortexed and sonicated with methanol and potassium  
169 hydroxide solution at 0°C. Subsequently, after incuba-  
170 tion at 85°C for 30 min, followed by cooling at 4°C for  
171 20 min, high-performance liquid chromatography  
172 (HPLC)-grade pentane was used to extract ergosterol  
173 from the soil mixture. Three pentane extracts were  
174 combined and dried under N<sub>2</sub> gas before being redis-  
175 solved in methanol and filtered through 0.2-µm Teflon  
176 filters, then analyzed on a Waters Sugar Analyzer I HPLC  
177 system (Elstree, Hertfordshire, UK). Incorporation of an  
178 internal ergosterol standard allowed the extraction effi-  
179 ciency of ergosterol from soil in these experiments to be  
180 estimated at  $87 \pm 4.7\%$ .

181 *Total Soil DNA Extraction and Purification.* Total  
182 soil DNA was extracted as previously described [8].  
183 Briefly, soil (0.5 g) was added to tubes containing glass and  
184 zirconia beads, to which hexadecyltrimethylammonium  
185 bromide (CTAB) extraction buffer was added. After  
186 incubation at 70°C for 10 min, phenol/chloroform/  
187 isoamylalcohol (25:24:1) was added and tubes were  
188 then shaken in a Hybaid Ribolyser (Ashford, UK) at  
189 5.5 m/s for 30 s. Following bead beating, tubes were  
190 centrifuged and the aqueous layer was removed and  
191 extracted twice with chloroform/isoamylalcohol (24:1).  
192 A further purification procedure was performed involving  
193 incubation with lysozyme solution (100 mg/mL) for 30  
194 min at 37°C. Tubes were again centrifuged and the  
195 aqueous layer removed and further purified using a High

Pure PCR Product Clean Up Kit (Roche Diagnostics 196  
GmbH, Penzberg, Germany) according to manufac- 197  
turer's instructions. DNA was eluted in 50 µl of the 198  
manufacturer's proprietary elution buffer solution and 199  
was consistently suitable for polymerase chain reac- 200  
tion (PCR) amplification without further treatment. 201

*Bacterial Community Fingerprinting by T-RFLP* 202  
*Analysis.* Terminal restriction ribotype lengths were 203  
determined using a modification of the method of Liu 204  
*et al.* [38]. After extraction and purification of total DNA 205  
from soil, the bacterial 16S small subunit rRNA gene was 206  
amplified using primer set F27 (5'-AGAGTTTGATCMT 207  
GGCTCAG-3') and R1492 (5'-TACGGYTACCTTGT 208  
TACGACT-3') [37]. The forward primer F27 was 209  
labeled with Beckman Coulter fluorescent dye D4 210  
(Invitrogen, Paisley, Scotland, UK). PCR reactions were 211  
performed in 50-µl volumes containing 5 µl of 10× Mg- 212  
free PCR buffer, 1.25 mM MgCl<sub>2</sub>, 15 pmol of each 213  
primer, 200 µM of each dNTP, 25 µg BSA, ~10 ng 214  
extracted total soil DNA, and 2.5 U *Taq* DNA polymerase 215  
(Promega, Southampton, UK). The thermocycling 216  
conditions were as follows: a hot start at 94°C for 3 min 217  
(1 cycle); 94°C for 1 min, 53°C for 2 min, 72°C for 2 min 218  
(26 cycles); 72°C for 7 min. PCR products were first 219  
visualized on a 1% agarose gel and purified using a High 220  
Pure PCR product purification kit (Roche Diagnostics) 221  
according to the manufacturer's instructions. The purified 222  
PCR product was quantified on a 1.2% agarose gel then 223  
digested enzymatically as follows: approximately 50 ng of 224  
PCR product was added to a reaction mixture containing 225  
sterile Millipore water, 20 U of restriction endonuclease 226  
*MspI*, and 2 µl of corresponding enzyme buffer. Digests 227  
were performed in a final volume of 20 µl and incubated 228  
in a water bath at 37°C for 4 h. Digests were desalted and 229  
aliquots (1 µl) were mixed with 38.75 µl of deionized 230  
formamide and 0.25 µl of Beckman Coulter size standard 231  
600 (High Wycombe, Bucks, UK). 232

*Fungal Community Fingerprinting by ARISA* 233  
*(Automated Ribosomal Intergenic Spacer Analysis).* 234  
The fungal genetic region containing the two internal 235  
transcribed spacers (ITS) and the 5.8S rRNA gene (ITS1– 236  
5.8S–ITS2) was amplified using primer set ITS1-F 237  
(CTTGGTCATTTAGAGGAAGTAA) [16] and ITS4 238  
(TCCTCCGCTTATTGATATGC) [62]. Amplified 239  
sequences contained the two ITS regions and the 5.8S 240  
gene plus 22 bp from the forward primer. The forward 241  
primer ITS1-F was labeled with Beckman Coulter 242  
fluorescent dye D4 (Invitrogen). PCR reactions were 243  
performed in 50-µl volumes of 25 µl master mix 244  
[containing 50 units/mL *Taq* DNA polymerase in a 245  
proprietary reaction buffer (pH 8.5), 400 µM each dNTP, 246  
and 3 mM MgCl<sub>2</sub>; Promega], 15 pmol each primer, 25 247  
µg BSA, and ~10 ng extracted total soil DNA. The 248

249 thermocycling conditions were as follows: a hot start at  
250 95°C for 4 min (1 cycle); 95°C for 1 min, 53°C for 30 s,  
251 72°C for 1 min (35 cycles); 72°C for 7 min (1 cycle). PCR  
252 products were first visualized on a 1.2% agarose gel, then  
253 purified using a High Pure PCR product purification kit  
254 (Roche Diagnostics) and eluted in 50 µl of sterile water  
255 heated to 50°C. The purified PCR products were  
256 quantified on a 1.2% agarose gel before mixing aliquots  
257 (0.5–1 µl, equivalent to ~2 ng DNA) with 38.4 µl of  
258 deionized formamide, 0.2 µl of Beckman Coulter size  
259 standard 600 (dye D1), and 0.4 µl of custom-made  
260 marker (containing ribotypes ranging from 600 to 1000  
261 bp in intervals of 20 bp, and 1000–1200 in intervals of  
262 50 bp, all labeled with Beckman Coulter Dye D1)  
263 (BioVentures, Murfreesboro, TN, USA).

264 **TRFLP and ARISA Ribotype Analysis.** Bacterial  
265 terminal restriction and fungal ARISA ribotype lengths were  
266 determined by electrophoresis using a Beckman Coulter  
267 (CEQ8000) automated sequencer, version 6.0.2 (High  
268 Wycombe, Bucks, UK) (resolution ±1 bp up to 400 bp  
269 ribotype length, ±2 bp thereafter). Run conditions were  
270 60°C separation temperature, 4 kV voltage, and 120 min  
271 separation time to allow for separation of the larger  
272 ribotypes. Analysis of ribotype profiles was performed  
273 using the Beckman Coulter ribotype analysis package 8000,  
274 version 8.0.52, ribotype analysis algorithm version 2.2.1  
275 (High Wycombe, Bucks, UK). A quartic polynomial model,  
276 rather than the recommended cubic model, was used for size  
277 standard calibration as this resulted in improved correlation  
278 between expected and actual size standard ribotype sizes,  
279 particularly for ribotypes in the range 400–1200 bp.  
280 Ribotypes with peak heights less than 1% of the total peak  
281 height for all ribotypes in a sample were regarded as  
282 background noise and excluded from analysis. Ribotypes  
283 that differed by less than 0.5 bp in different profiles were  
284 considered identical. The peak heights of individual ribo-  
285 types were relativized as a percentage of their abundance  
286 within a sample to account for DNA quantity differences  
287 between replicate profiles, resulting in profiles containing  
288 data on ribotypes present and their relative abundances.

289 **Statistical Analysis.** Results for chemical analyses,  
290 microbial activity, fungal biomass, ribotype number, and  
291 individual abundances of the top 20 overall most  
292 abundant bacterial and fungal ribotypes were analyzed  
293 by one-way factorial analysis of variance (ANOVA) using  
294 Genstat v6 (VSN International, Oxford, UK). The signifi-  
295 cance level was set at  $p < 0.05$ . In addition, Genstat was  
296 used to calculate the overall abundance of each individ-  
297 ual ribotype for all samples. Ribotypes were then ranked  
298 according to overall abundance and these rankings were  
299 used to select individual ribotypes for ANOVA. Bacterial  
300 TRFLP and fungal ARISA profiles were analyzed using  
301 Canoco for Windows, v4.02 (Centre for Biometry,

Wageningen, The Netherlands). Initial analysis by 302  
detrended correspondence analysis (DCA) revealed that 303  
bacterial TRFLP data exhibited a linear response to the 304  
environmental variables (grassland type and site), there- 305  
by indicating redundancy analysis (RDA) as the most 306  
appropriate multivariate approach, whereas fungal ARISA 307  
data exhibited a unimodal response, indicating canonical 308  
correspondence analysis (CCA) was most appropriate. 309  
Initially, RDA or CCA of all ribotypes with all environ- 310  
mental factors, including all of their interactions, was 311  
carried out. This resulted in overly complex analysis results 312  
and diagrams, necessitating a reduction in the amount of 313  
data analyzed. The environmental factors most important 314  
in explaining variation in ribotype profiles were selected by 315  
eliminating those with low canonical coefficients and  $t$ - 316  
values [33]. Next, analysis was limited to the top 20 317  
ribotypes, as ranked by abundance. As there was little or no 318  
change in analysis results after limiting the environmental 319  
variables and ribotype set, it was concluded that the 320  
variables and ribotypes accounting for the majority of 321  
variance in the data had been selected. The resulting 322  
ordination biplots approximated the weighted average of 323  
each species (in this case, relative abundances of ribotypes) 324  
with respect to each of the environmental variables, which 325  
were represented as arrows. The length of these arrows 326  
indicated the relative importance of that environmental 327  
factor in explaining variation in ribotype profiles, whereas 328  
the angle between arrows indicated the degree to which 329  
they were correlated [33]. A Monte Carlo permutation test 330  
based on 199 random permutations was used to test the 331  
null hypothesis that ribotype profiles were unrelated to 332  
environmental variables. 333

## 334 Results

**Grassland Floristic Composition.** Vegetational analysis 335  
revealed that unimproved U4a grasslands were domi- 336  
nated by *A. capillaris*, *Anthoxanthum odoratum*, and 337  
*F. ovina*, with high frequencies of *Potentilla erecta* and 338  
*G. saxatile*. Semi-improved U4b grasslands, although still 339  
dominated by *A. capillaris*, showed reduced abundances 340  
of *F. ovina* and increased occurrences of *F. rubra*. 341  
Additionally, more mesophytic species such as *Holcus* 342  
*lanatus* and *T. repens* were evident. 343

**Soil Chemical Composition.** Analysis of soil pH, 344  
nitrogen, phosphorus, and potassium (Table 1) indicated 345  
soil chemical composition varied due to both grassland 346  
type and site. On average, over all sites, U4a grassland 347  
soils had a slightly lower pH, higher nitrogen content, 348  
and lower phosphorus and potassium content than U4b 349  
grassland soils. However, these trends varied by site. 350

**Microbial Activity and Fungal Biomass.** To 351  
determine broad-scale trends in soil microbial community 352

353 structure, microbial activity and fungal biomass were  
 354 determined (Table 2). Microbial activity varied widely  
 355 between sites, with soils from Kings River having the  
 356 highest values, whereas soils from Annagh Hill and Sally  
 357 Gap had the lowest. Grassland type had a smaller but  
 358 significant impact, with microbial activity higher in U4b  
 359 grassland soils than in U4a grassland soils at every site.  
 360 Fungal biomass, measured as ergosterol, also experienced  
 361 wide variation due to site, with soils from Long Hill  
 362 highest in fungal biomass, while those from Sally Gap  
 363 were lowest (Table 2). Grassland type and site interacted  
 364 significantly, with soil from U4b grasslands lower in  
 365 fungal biomass than those from U4a grassland, at every  
 366 site except Kings River.

367 **Bacterial and Fungal Ribotype Number.** Bacterial  
 368 TRFLP detected a total of 89 unique terminal fragments  
 369 (ribotypes) after analysis of all samples, ranging in size  
 370 from 60 to 640 bp. A total of 453 fungal ribotypes was  
 371 detected by fungal ARISA, ranging in size from 47 to  
 372 1100 bp. Bacterial ribotype number (Table 2) was  
 373 affected by a significant interaction between grassland  
 374 type and site, with numbers higher in soils from U4b  
 375 grasslands than in U4a grassland soil at three sites, but  
 376 the reverse true at the other two sites. Fungal ribotype  
 377 number (Table 2) was significantly affected only by  
 378 grassland type, with fungal ribotype numbers higher in  
 379 U4a grassland soil than in U4b soil at every site.

380 **Bacterial Community Structure.** The effects of  
 381 grassland type and site on bacterial community structure  
 382 were further explored using redundancy analysis (RDA),  
 383 with results shown in the ordination plot in Fig. 1. The top  
 384 20 most abundant bacterial ribotypes, accounting for 86.4%  
 385 of total abundance, were included in the analysis. Axes 1

and 2 together accounted for 30.8% of total variation 386  
 within bacterial ribotype profiles, and 72.6% of variation in 387  
 profiles that could be attributed to the environmental 388  
 variables of grassland type and site. Species–environment 389  
 correlations for both axes were quite high (over 0.81), 390  
 indicating that changes in bacterial ribotype profiles were 391  
 closely correlated with changes in grassland type and site. 392  
 Both the first axis on its own, then all axes together, 393  
 were analyzed by the Monte Carlo test and found to 394  
 explain a significant ( $p < 0.05$ ) amount of variation within 395  
 the data. 396

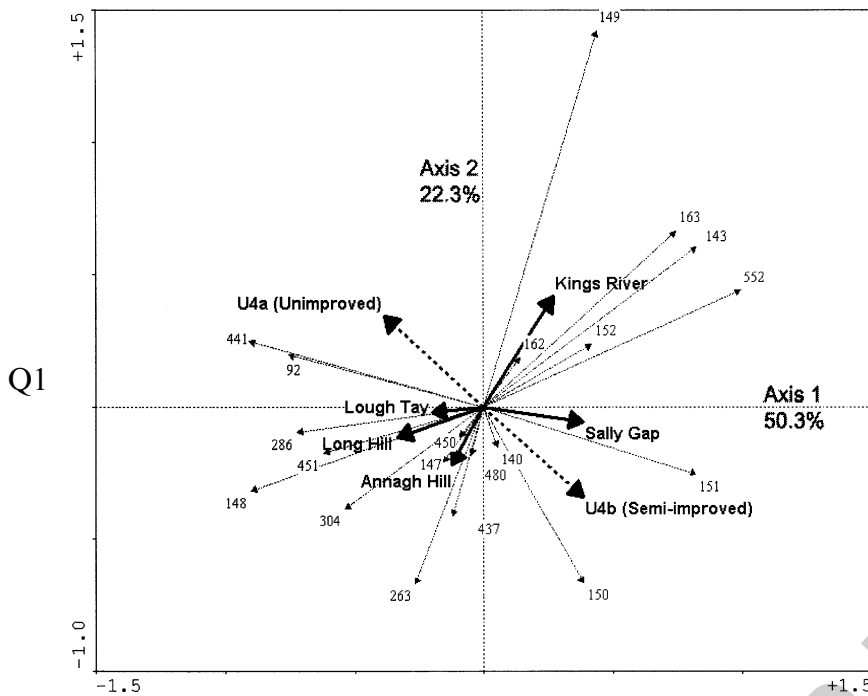
Intraset correlations revealed that Axis 1 was most 397  
 closely correlated with grassland type (U4a and U4b), 398  
 indicating that grassland type had the biggest impact on 399  
 community structure, while Axis 2 was most closely 400  
 correlated with the Sally Gap site. These observations 401  
 were supported by noting the length of the arrows for 402  
 U4a and U4b grassland types and the Sally Gap site in the 403  
 RDA plot in Fig. 1, which appeared longer than arrows 404  
 for other sites, indicating that these factors had a large 405  
 impact on bacterial community structure. The arrow for 406  
 the Kings River site was also long, and values for its 407  
 intraset correlations confirmed that it also had a large 408  
 effect on bacterial community profiles. 409

Correlations shown in the RDA plot were supported 410  
 by ANOVA results (Table 3); for example, the arrow for 411  
 the second most abundant ribotype, TRF 150, appeared 412  
 close to the arrow for U4b grassland type on the RDA 413  
 plot, and this relationship was confirmed by ANOVA, 414  
 which found that the abundance of TRF 150 was 415  
 significantly affected by grassland type, and that its 416  
 abundance was highest in U4b grassland soil. Grassland 417  
 type was found to have a significant effect on many 418  
 (50%) of the top 20 most abundant ribotypes. Site also 419  
 had a strong effect on abundances of the top 20 420

t2.1 **Table 2. Microbial activity, fungal biomass, and bacterial and fungal ribotype numbers of soils from U4a and U4b grassland types at each field site ( $n = 3$ ), and the average value for each grassland type over all sites ( $n = 15$ )**

t2.2	Microbial activity ( $\mu\text{g TPF g}^{-1}$ dry soil)		Fungal Biomass ( $\mu\text{g ergosterol g}^{-1}$ soil)		Bacterial ribotype number (mean TRFs per replicate)		Fungal ribotype number (mean fragments per replicate)		
	U4a	U4b	U4a	U4b	U4a	U4b	U4a	U4b	
t2.3									
t2.4	Site								
t2.5	Long Hill	5	190	3.27	1.43	14	31	38	26
t2.6	Sally Gap	9	52	0.05	0.01	19	31	28	26
t2.7	Lough Tay	31	87	2.01	0.25	6	10	40	18
t2.8	Kings River	388	520	0.44	2.24	17	9	29	25
t2.9	Annagh Hill	2	20	1.92	1.10	27	9	35	16
t2.10	Average	86	174	1.53	1.00	17	18	34	22
t2.11	SED (Site*Grassland type)	69.4		0.842		8.3		12.1	
t2.12	$p$ Values								
t2.13	Site	***		*		*		NS	
t2.14	Grassland type	*		NS		NS		*	
t2.15	Site*Grassland type	NS		*		*		NS	

t2.16 Means and standard error of differences (SED) are shown. ANOVA  $p$  values for site, grassland type and their interaction are shown as NS, not significant; \* $p < 0.05$ ; \*\*\* $p < 0.001$ .



**Figure 1.** Redundancy analysis (RDA) ordination diagram of bacterial TRFLP data, with grassland type (*dashed lines*) and sites (*bold solid lines*) represented as *large arrows* and bacterial TRF ribotypes represented as *smaller arrows*. TRFs are labeled according to ribotype size (bp). Axis 1 explains 50.3% of the TRFLP–environment variance, whereas Axis 2 explains a further 22.3% of the TRFLP–environment variance.

429 ribotypes, having a significant effect on 30% of them. 430 Ribotypes that were significantly affected by site often had 431 highest abundances at Sally Gap or Kings River, indicating 432 these sites significantly affected the abundances of dom- 433 inant ribotypes. In addition, significant interactions 434 between grassland type and site were also noted.

**Fungal Community Structure.** Fungal ARISA 435 ribotype profiles were investigated using canonical corre- 436 spondence analysis (CCA) after initial data exploration 437 revealed fungal community profiles responded to 438 environmental variables (grassland type and site) in a 439 unimodal fashion. The top 20 most abundant fungal 440

t3.1 **Table 3.** Abundance rankings and ANOVA results for the top 20 most abundant bacterial TRFLP fragments, as ranked by average abundance over all samples

TRFLP fragment (bp)	Abundance			p Values	
	Rank	% Abundance	Cumulative % abundance	Grassland type	Site
92	1	23.4	23.4	0.084	<b>0.015</b> (LT)
150	2	12.6	36.0	<b>0.007</b> (U4b)	0.273
151	3	11.8	47.8	0.074	<b>0.011</b> (SG)
148	4	7.8	55.6	<b>0.039</b> (U4a)	0.100
441	5	4.2	59.8	<b>0.004</b> (U4a) <sup>a</sup>	0.051 <sup>a</sup>
286	6	3.3	63.1	<b>0.043</b> (U4a)	0.209
143	7	2.8	65.9	0.705	< <b>0.001</b> (SG)
552	8	2.6	68.5	<b>0.011</b> (U4b) <sup>a</sup>	< <b>0.001</b> (KR) <sup>a</sup>
152	9	2.3	70.8	0.604	0.341
149	10	2.3	73.1	<b>0.003</b> (U4a) <sup>a</sup>	< <b>0.001</b> (KR) <sup>a</sup>
263	11	2.2	75.3	<b>0.038</b> (U4a)	0.339(AH)
304	12	1.8	77.1	<b>0.007</b> (U4a)	0.654
163	13	1.4	78.5	<b>0.035</b> (U4b)	0.654
480	14	1.4	80.0	0.809	0.916
451	15	1.4	81.3	<b>0.028</b> (U4a)	0.158
147	16	1.1	82.4	0.265	0.670
437	17	1.1	83.5	0.358	0.499
140	18	1.1	84.5	0.917	0.657
450	19	0.9	85.5	0.058 <sup>a</sup>	0.823 <sup>a</sup>
162	20	0.9	86.4	0.492	0.431

t3.24 Data in bold indicate a significant effect ( $p < 0.05$ ).

Grassland type/site with highest abundance of fragment is indicated next to significant effect (LT: Lough Tay; SG: Sally Gap; KR: Kings River; AH: Annagh Hill).

<sup>a</sup>Significant interaction with each other.

441 ribotypes, accounting for 38.4% of total abundance, were  
 442 included in the analysis. Results of CCA (Fig. 2) showed  
 443 both Axes 1 and 2 had high eigenvalues (0.556 and  
 444 0.303), indicating that grassland type and site accounted  
 445 for a large percentage of variation between fungal  
 446 profiles. This was confirmed by calculations that  
 447 indicated Axes 1 and 2 together accounted for 17.0% of  
 448 the total variation within the data, and 63.6% of the  
 449 variation that could be attributed to grassland type and  
 450 site. Species–environment correlations were also very  
 451 high (above 0.76), indicating that changes in fungal  
 452 community profiles corresponded strongly with changes  
 453 in grassland type and site. Monte Carlo tests of significance  
 454 for the first axis alone, and all axes combined, indicated the  
 455 analysis accounted for a significant percentage of variation  
 456 within fungal ARISA data ( $p < 0.05$ ).

457 The longest arrows in the CCA ordination diagram  
 458 (Fig. 2) were for the sites Annagh Hill and Kings River,  
 459 indicating these had the greatest impact on community  
 460 structure. This was confirmed by intraset correlations,  
 461 which revealed that Axis 1 was most strongly correlated  
 462 with Annagh Hill, while Axis 2 was associated with Kings  
 463 River. Grassland type (U4a and U4b) also had long  
 464 arrows and high intraset correlations, suggesting that  
 465 grassland type was also an important factor affecting  
 466 fungal community structure. A third site, Sally Gap, also  
 467 had a long arrow in the CCA plot and high correlation

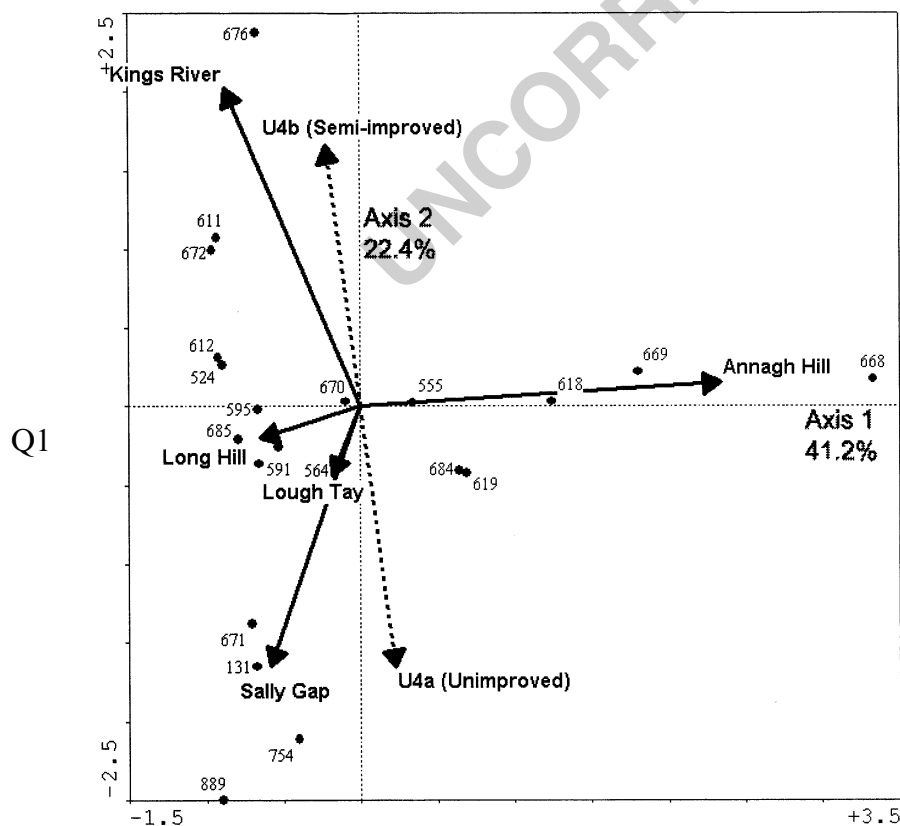
values, indicating it affected fungal community structure. 468  
 In contrast, arrows for the Lough Tay and Long Hill sites 469  
 were notably shorter than those for the other sites and 470  
 grassland type, signifying that these had a relatively 471  
 smaller impact on fungal communities. 472

These observations were supported by ANOVA of the 473  
 top 20 most abundant ribotypes (Table 4), which revealed 474  
 that grassland type had a significant effect on abundances 475  
 of four of the top 20 ribotypes. Interestingly, all of the four 476  
 affected had highest abundances in U4b grassland soil. 477  
 The effect of site was also evident, with those ribotypes 478  
 significantly affected by site having the highest abundances 479  
 in Annagh Hill, Kings River, and Sally Gap. In addition, 480  
 significant interactions between grassland type and site 481  
 affected the abundances of several fungal ribotypes. 482

## Discussion

483

The effect of a moderate change in grassland floristic 484  
 composition (from unimproved U4a to semi-improved 485  
 U4b) on microbial community structure was investigated 486  
 at five geographically separate field sites with similar 487  
 underlying geological substrata. Both grassland type and 488  
 site were found to impact microbial community struc- 489  
 ture, but these effects varied and neither influence had a 490  
 consistently larger impact than the other. The original 491



**Figure 2.** Canonical correspondence analysis (CCA) ordination diagram of fungal ARISA data, with grassland type (dashed lines) and sites (bold solid lines) represented as large arrows and fungal ARISA ribotypes represented as black dots. Ribotypes are labeled according to ribotype size (bp). Axis 1 explains 41.2% of the FARISA–environment variance, whereas Axis 2 explains a further 22.4% of the FARISA–environment variance.

t4.1 **Table 4. Abundance rankings and ANOVA results for the top 20 most abundant fungal ARISA fragments, as ranked by average abundance over all samples**

t4.2		Abundance			p values	
t4.3	ARISA fragment (bp)	Abundance rank	% Abundance	Cumulative % abundance	Grassland type	Site
t4.4	668	1	3.6	3.6	0.754	0.073
t4.5	670	2	3.0	6.6	0.624 <sup>a</sup>	0.196 <sup>a</sup>
t4.6	524	3	2.5	9.1	<b>0.049</b> (U4b) <sup>a</sup>	<b>0.047</b> (SG) <sup>a</sup>
t4.7	555	4	2.5	11.6	0.485	0.458
t4.8	684	5	2.4	14.0	0.527	0.326
t4.9	671	6	2.3	16.3	0.859	0.358
t4.10	595	7	2.2	18.5	0.848	0.417
t4.11	672	8	2.1	20.7	<b>0.015</b> (U4b) <sup>a</sup>	<b>0.001</b> (KR) <sup>a</sup>
t4.12	611	9	2.0	22.7	<b>0.005</b> (U4b)	<b>0.025</b> (KR)
t4.13	612	10	1.9	24.6	0.137	0.131
t4.14	669	11	1.7	26.3	0.264	0.163
t4.15	131	12	1.6	27.9	0.556	0.771
t4.16	564	13	1.4	29.4	0.060	0.601
t4.17	685	14	1.4	30.8	<b>0.033</b> (U4b)	0.099
t4.18	619	15	1.4	32.2	0.072	0.826
t4.19	618	16	1.3	33.5	0.081	<b>0.049</b> (AH)
t4.20	889	17	1.3	34.9	0.431	0.329
t4.21	676	18	1.3	36.2	0.193	0.063
t4.22	754	19	1.1	37.3	0.498	0.492
t4.23	591	20	1.1	38.4	0.533	0.061

t4.24 Data in bold indicate a significant effect ( $p < 0.05$ ).

Grassland type/site with highest abundance of fragment is indicated next to significant effect (SG: Sally Gap; KR: Kings River; AH: Annagh Hill).

<sup>a</sup>Significant interaction with each other.

493 question posed in this study was whether a grassland  
494 classification system such as the UK National Vegetation  
495 Classification (NVC) could be used as a predictor of mi-  
496 crobial community structures at geographically separated  
497 sites with similar geological origins. Our results indicate  
498 that this hypothesis must be rejected, as microbial  
499 community measurements varied widely according to  
500 site, and trends in the effect of grassland type were  
501 inconsistent.

502 Strong variation in soil physicochemical properties  
503 between sites was noted, indicating that grassland type  
504 did not correspond with soil properties. Changes in pH  
505 due to semi-improvement were evident, with U4a grass-  
506 lands having a slightly lower pH overall, although this  
507 effect varied between sites. Soil nitrogen levels also varied  
508 between sites, as has been noted elsewhere in geograph-  
509 ically separated sites with similar grassland types [23].  
510 Although some site variation was noted in soil phospho-  
511 rus and potassium, at most sites U4b grassland soil had  
512 higher phosphorus and potassium contents than U4a,  
513 possibly due either to movement of these elements from  
514 improved areas, or from increased nutrient mobilization  
515 due to increased pH [6].

516 Although microbial activity was strongly affected by  
517 site, there was a consistent influence of grassland type,  
518 with semi-improved U4b soils having higher microbial  
519 activity than unimproved U4a soils from the same site.  
520 This increase in activity from unimproved to semi-

improved soils was also noted by Brodie *et al.* [8] and 521  
Williams *et al.* [63]. The decrease seen in fungal biomass 522  
in semi-improved soils at most sites indicates that that 523  
there may be more fungi in unimproved soils, which has 524  
been a common finding in studies of upland grasslands 525  
[1, 2, 4, 9, 23, 24]. 526

Bacterial ribotype numbers were significantly affected 527  
by site, but increased from unimproved U4a to semi- 528  
improved U4b grassland soil at most sites, mirroring 529  
trends in activity. An increase in bacterial ribotype 530  
number from unimproved to semi-improved soils was 531  
also noted by Brodie *et al.* [8]. Bacterial community 532  
structure was explored by multivariate analysis, which 533  
indicated that although grassland type had the largest 534  
influence, the Sally Gap and Kings River sites also had a 535  
strong impact. The large impact of unimproved and 536  
semi-improved grassland types on bacterial community 537  
structure indicated that semi-improvement resulted in 538  
differing bacterial community structures. Other studies 539  
have also observed changes in bacterial community 540  
structure after semi-improvement [8, 43]. The strong 541  
impact of the Kings River and Sally Gap sites may be 542  
related to high microbial activity and low soil nutrient 543  
levels, respectively. 544

Fungal ribotype number was affected significantly by 545  
grassland type, with fungal ribotypes decreasing in semi- 546  
improved U4b soil as compared to unimproved U4a soil 547  
at all sites, which corresponded with the decrease in 548



549 fungal biomass noticed at most sites. Although site did  
550 not significantly affect fungal ribotype number, canonical  
551 correspondence analysis (CCA) of fungal ARISA profiles  
552 revealed that certain sites (Annagh Hill, Kings River, and  
553 Sally Gap) had a large influence on fungal community  
554 structure. Grassland type also had an important influence  
555 on soil fungal populations, with unimproved U4a and  
556 semi-improved U4b soils having differing fungal com-  
557 munity structures.

558 Studies of geographically separated upland acidic  
559 grasslands have been conducted previously. Clegg *et al.*  
560 [12], studying unimproved (U4a), semi-improved (U4b),  
561 and improved (MG6) grassland soils at three sites, found  
562 that complexity rankings of DNA from different grass-  
563 land types varied between sites, based upon percent G +  
564 C content. Grayston *et al.* [23, 24], working on similar  
565 sites and grasslands, found differences between grassland  
566 types at each site, with variation between sites and  
567 seasons, using community-level physiological profiling  
568 (CLPP), phospholipid fatty acid contents (PLFA), and  
569 percent G + C profiling. These studies, in concurrence  
570 with the results presented here, indicate a large impact  
571 of site on microbial community structure, often obscur-  
572 ing differences between grassland types. However, they  
573 have relied upon broad-scale analyses such as percent G  
574 + C content, PLFA, and CLPP. The results presented in  
575 this work mark the first time that DNA-based profiling  
576 approaches (TRFLP and ARISA) have been applied  
577 together to a geographic survey of upland acidic grass-  
578 lands. DNA-based community fingerprinting approaches  
579 can be difficult to interpret because of species differences  
580 in rRNA gene copy number [13, 15], biases resulting  
581 from PCR amplification [34, 47, 56], and the difficulty  
582 in standardizing the amount of DNA analyzed in each  
583 replicate [14, 39]. However, as all samples were subject to  
584 the same biases, it was still possible to compare them on a  
585 relative basis, especially after standardization of ribotype  
586 peak heights into proportions per sample.

587 There are several possible reasons for the lack of  
588 grouping in microbial communities in accordance with  
589 grassland type. It may be that changes in floristic  
590 composition due to semi-improvement did not signifi-  
591 cantly affect microbial community structure, or were too  
592 small to have consistent significant effects. There is  
593 evidence suggesting that certain plant species impact  
594 microbial community structure more than others [32,  
595 42]. Other studies on the impact of plant community  
596 composition on soil microbial communities [10, 18, 25,  
597 26, 29, 31, 59] indicate that changes in functional groups  
598 of plants may have stronger impacts than plant diversity  
599 or composition *per se*. In our study, floristic composition  
600 changed from *Festuca–Agrostis–Galium* domination in  
601 unimproved (U4a) grasslands, to *Festuca–Agrostis–*  
602 *Galium* with *Holcus–Trifolium* subcommunities in semi-  
603 improved (U4b) grasslands. Although *Trifolium* is a

legume, both grassland types were dominated by grasses. 604  
Therefore, there was little change in plant functional 605  
group composition, which may have contributed to the 606  
inconsistency of grassland type effects on microbial 607  
community structure. 608

Site can have a considerable impact on microbial 609  
community structure [12]. Several studies have found 610  
strong links between soil physicochemical properties 611  
such as soil texture and type, and microbial community 612  
structure [5, 19, 25, 32, 51, 58]. Correlations between 613  
soil physicochemistry and microbial community struc- 614  
ture have been found both in field and microcosm 615  
studies on acidic upland grassland soils [8, 9, 24, 30, 35]. 616  
In our study, site affected all soil physicochemical 617  
characteristics (pH, percent N, phosphorus, potassium) 618  
significantly, and effects of semi-improvement on soil 619  
chemistry also varied significantly between sites. These 620  
inconsistent trends between sites may mean that minor 621  
changes in soil chemistry within a site could be sufficient 622  
to support a change in floristic composition, but are 623  
not enough to strongly affect soil microbial popula- 624  
tions. Often, changes in soil chemical parameters were 625  
greater between sites than between grassland types at 626  
a site, which may contribute to the impact of site on 627  
microbial community structure. For example, the same 628  
floristic composition can occur at different sites at a range 629  
of pH values (published values for U4a grassland soils 630  
range from 3.3 to 6.3, and for U4b grassland soils range 631  
from 3.98 to 6.4 [8, 11, 12, 23, 24, 60]); such large 632  
differences in pH between sites with the same floristic 633  
grassland classification could strongly affect soil micro- 634  
bial communities. 635

In practice, the real determinants of microbial com- 636  
munity structure between sites and grassland types are 637  
likely to be complex interactions between plant composi- 638  
tion effects, and individual site characteristics, including 639  
soil physicochemical composition. A recent study of an 640  
unimproved (U4a) grassland in Scotland reported that 641  
CLPP, DGGE, and PLFA profiles were affected by both 642  
vegetation class and soil physical and chemical factors, but 643  
no single factor or factors stood out [48]; rather, it 644  
appeared that these interacted in complex mechanisms 645  
to influence soil microbial communities. It is likely that a 646  
similar situation exists in relation to the sites and 647  
grassland types investigated in our study. 648

### Acknowledgments 649

Part of this work was supported by an Environmental 650  
Protection Agency contributory scholarship under the 651  
Environmental Research Technological Development and 652  
Innovation (ERTDI) program. We thank Dr. John 653  
Connolly of Statistics Department, UCD for statistical 654  
advice and guidance, and John Flynn for technical 655  
assistance. 656

657 **References**

- 658 1. Bardgett, RD, Hobbs, PJ, Frostegård, A (1996) Changes in soil  
659 fungal:bacterial biomass ratios following reductions in the intensity  
660 of management of an upland grassland. *Biol Fertil Soils* 22: 261–264
- 661 2. Bardgett, RD, Lovell, RD, Hobbs, PJ, Jarvis, SC (1999) Seasonal  
662 changes in soil microbial communities along a fertility gradient of  
663 temperate grasslands. *Soil Biol Biochem* 31: 1021–1030
- 664 3. Bardgett, RD, Mawdsley, JL, Edwards, S, Hobbs, PJ, Rodwell, JS,  
665 Davies, WJ (1999) Plant species and nitrogen effects on soil  
666 biological properties of temperate upland grasslands. *Funct Ecol*  
667 13: 650–660
- 668 4. Bardgett, RD, McAlister, E (1999) The measurement of soil  
669 fungal:bacterial biomass ratios as an indicator of ecosystem self-  
670 regulation in temperate meadow grasslands. *Biol Fertil Soils* 29:  
671 282–290
- 672 5. Bossio, DA, Scow, KM, Gunapala, N, Graham, KJ (1998)  
673 Determinants of soil microbial communities: effects of agricultural  
674 management, season, and soil type on phospholipid fatty acid  
675 profiles. *Microb Ecol* 36: 1–12
- 676 6. Brady, NC, Weil, RR (2002) *The Nature and Properties of Soils*.  
677 Prentice-Hall, Upper Saddle River, NJ, USA, pp 592–637
- 678 7. Bremner, JM (1996) Nitrogen—total. In: Sparks, DL, Page, AL,  
679 Helmke, PA, Loepfert, RH, Soltanpour, PN, Tabatabai, MA,  
680 Johnston, CT, Sumner, ME (Eds.) *Methods of Soil Analysis: Part*  
681 *3, Chemical Methods*. Soil Science Society of America, Madison,  
682 WI, pp 1085–1121
- 683 8. Brodie, E, Edwards, S, Clipson, N (2002) Bacterial community  
684 dynamics across a floristic gradient in a temperate upland  
685 grassland ecosystem. *Microb Ecol* 44: 260–270
- 686 9. Brodie, E, Edwards, S, Clipson, N (2003) Soil fungal community  
687 structure in a temperate upland grassland soil. *FEMS Microbiol*  
688 *Ecol* 45: 105–114
- 689 10. Buckley, DH, Schmidt, TM (2003) Diversity and dynamics of  
690 microbial communities in soils from agro-ecosystems. *Environ*  
691 *Microbiol* 5: 441–452
- 692 11. Clegg, CD, Ritz, K, Griffiths, BS (1998) Broad-scale analysis of soil  
693 microbial community DNA from upland grasslands. *Antonie van*  
694 *Leeuwenhoek* 73: 9–14
- 695 12. Clegg, CD, Ritz, K, Griffiths, BS (2000) %G + C profiling and  
696 cross-hybridisation of microbial DNA reveals great variation in  
697 below-ground community structure in UK upland grasslands.  
698 *Appl Soil Ecol* 14: 125–134
- 699 13. Crosby, LD, Criddle, CS (2003) Understanding bias in microbial  
700 community analysis techniques due to *rnn* operon copy number  
701 heterogeneity. *BioTechniques* 34: 790–802
- 702 14. Egert, M, Friedrich, MW (2003) Formation of pseudo-terminal  
703 restriction ribotypes, a PCR-related bias affecting terminal restric-  
704 tion length polymorphism analysis of microbial community  
705 structure. *Appl Environ Microbiol* 69: 2555–2562
- 706 15. Fogel, GB, Collins, CR, Li, J, Brunk, CF (1999) Prokaryotic  
707 genome size and SSU rDNA copy number: estimation of microbial  
708 relative abundance from a mixed population. *Microb Ecol* 38:  
709 93–113
- 710 16. Gardes, M, Bruns, TD (1993) ITS primers with enhanced  
711 specificity for basidiomycetes—application to the identification  
712 of mycorrhizae and rusts. *Mol Ecol* 2: 113–118
- 713 17. Gardiner MJ, Radford T (1980) Soil survey bulletin number 36:  
714 Soil associations of Ireland and their land use potential. National  
715 Soil Survey of Ireland, An Foras Talúntais (The Agricultural  
716 Institute), Dublin, Ireland
- 717 18. Gastine, A, Scherer-Lorenzen, M, Leadley, PW (2003) No  
718 consistent effects of plant biodiversity on root biomass, soil biota  
719 and soil abiotic conditions in temperate grassland communities.  
720 *Appl Soil Ecol* 24: 101–111
19. Gelsomino, A, Keijzer-Wolters, AC, Cacco, G, van Elsas, JD (1999) 721  
Assessment of bacterial community structure in soil by polymer- 722  
ase chain reaction and denaturing gradient gel electrophoresis. 723  
*J Microbiol Methods* 38: 1–15 724
20. Geological Survey of Ireland (2000) *Geology of Kildare–Wicklow* 725  
(Sheet 16) 1:100,000. Geological Survey of Ireland. Ballsbridge, 726  
Dublin, Ireland 727
21. Gollotte, A, van Tuinen, D, Atkinson, D (2004) Diversity of 728  
arbuscular mycorrhizal fungi colonising roots of the grass species 729  
*Agrostis capillaris* and *Lolium perenne* in a field experiment. 730  
*Mycorrhiza* 14: 111–117 731
22. Grayston, SJ, Wang, S, Campbell, CD, Edwards, AC (1998) 732  
Selective influence of plant species on microbial diversity in the 733  
rhizosphere. *Soil Biol Biochem* 30: 369–378 734
23. Grayston, SJ, Griffith, GS, Mawdsley, JL, Campbell, CD, Bardgett, 735  
RD (2001) Accounting for variability in soil microbial communi- 736  
ties of temperate upland grassland ecosystems. *Soil Biol Biochem* 737  
33: 533–551 738
24. Grayston, SJ, Campbell, CD, Bardgett, RD, Mawdsley, JL, Clegg, 739  
CD, Ritz, K, Griffiths, BS, Rodwell, JS, Edwards, SJ, Davies, WJ, 740  
Elston, DJ, Millard, P (2004) Assessing shifts in microbial 741  
community structure across a range of grasslands of differing 742  
management intensity using CLPP, PLFA and community DNA 743  
techniques. *Appl Soil Ecol* 25: 63–84 744
25. Groffman, PM, Eagan, P, Sullivan, WM, Lemunyon, JL (1996) 745  
Grass species and soil type effects on microbial biomass and 746  
activity. *Plant Soil* 183: 61–67 747
26. Hedlund, K, Santa Regina, I, Van der Putten, WH, Lepš, J, Díaz, T, 748  
Korthals, GW, Lavorel, S, Brown, VK, Gormsen, D, Mortimer, SR, 749  
Rodríguez Barrueco, C, Roy, J, Smilauer, P, Smilauerová, M, Van 750  
Dijk, C (2003) Plant species diversity, plant biomass and responses 751  
of the soil community on abandoned land across Europe: 752  
idiosyncrasy or above-belowground time lags. *Oikos* 103: 45–58 753
27. Helmke, PA, Sparks, DL (1996) Lithium, sodium, potassium, 754  
rubidium, and cesium. In: Sparks, DL, Page, AL, Helmke, PA, 755  
Loepfert, RH, Soltanpour, PN, Tabatabai, MA, Johnston, CT, 756  
Sumner, ME (Eds.) *Methods of Soil Analysis: Part 3, Chemical* 757  
*Methods*. Soil Science Society of America, Madison, WI, pp 551–574 758
28. Hill, GT, Mitkowski, NA, Aldrich-Wolfe, L, Emele, LR, Jurkonie, 759  
DD, Ficke, A, Maldonado-Ramirez, S, Lynch, ST, Nelson, EB 760  
(2000) Methods for assessing the composition and diversity of soil 761  
microbial communities. *Appl Soil Ecol* 15: 25–36 762
29. Hooper, DU, Vitousek, PM (1997) The effects of plant composi- 763  
tion and diversity on ecosystem processes. *Science* 277: 1302–1305 764
30. Innes, L, Hobbs, PJ, Bardgett, RD (2004) The impacts of 765  
individual plant species on rhizosphere microbial communities 766  
in soils of different fertility. *Biol Fertil Soils* 40: 7–13 767
31. Johnson, D, Booth, RE, Whiteley, AS, Bailey, MJ, Read, DJ, Grime, 768  
JP, Leake, JR (2003) Plant community composition affects the 769  
biomass, activity and diversity of microorganisms in limestone 770  
grassland soil. *Eur J Soil Biol* 54: 671–677 771
32. Johnson, MJ, Lee, KY, Scow, KM (2003) DNA fingerprinting 772  
reveals links among agricultural crops, soil properties, and the 773  
composition of soil microbial communities. *Geoderma* 114: 774  
279–303 775
33. Jongman, RHG, ter Braak, CJF, van Tongeren, OFR (1995) *Data* 776  
*Analysis in Community and Landscape Ecology*. Cambridge 777  
University Press, Cambridge, pp 91–173 778
34. Jumpponen, A (2003) Soil fungal community assembly in a 779  
primary successional glacier forefront ecosystem as inferred from 780  
rDNA sequence analysis. *New Phytol* 158: 569–578 781
35. Kennedy, N, Brodie, E, Connolly, J, Clipson, N (2004) Impact of 782  
lime, nitrogen and plant species on bacterial community structure 783  
in grassland microcosms. *Environ Microbiol* 6: 1070–1080 784
36. Knudsen, D, Beegle, D (1988) Recommended phosphorus tests. 785

- 786 In: Dahnke WC (Ed.) Recommended Soil Tests Procedures for the  
787 North Central Region: Bulletin Number 499 (revised). North  
788 Dakota Agricultural Experiment Station. Fargo, ND, pp 12–15
- 789 37. Lane, DJ (1991) 16S/23S rRNA sequencing. In: Stackebrandt E,  
790 Goodfellow M (Eds.) Nucleic Acid Techniques in Bacterial  
791 Systematics. John Wiley & Sons, Chichester, pp 115–175
- 792 38. Liu, W-T, Marsh, TL, Cheng, H, Forney, LJ (1997) Characteriza-  
793 tion of microbial diversity by determining terminal restriction  
794 ribotype length polymorphisms of genes encoding 16S rRNA. Appl  
795 Environ Microbiol 63: 4516–4522
- 796 39. Lueders, T, Friedrich, MW (2003) Evaluation of PCR amplification  
797 bias by terminal restriction ribotype length polymorphism analysis  
798 of small-subunit rRNA and *mcrA* genes by using defined template  
799 mixtures of methanogenic pure cultures and soil DNA extracts.  
800 Appl Environ Microbiol 69: 320–326
- 801 40. Lynch, JM, Whipps, JM (1990) Substrate flow in the rhizosphere.  
802 Plant Soil 129: 1–10
- 803 41. Marilley, L, Aragno, M (1999) Phylogenetic diversity of bacterial  
804 communities differing in degree of proximity of *Lolium perenne*  
805 and *Trifolium repens* roots. Appl Soil Ecol 13: 127–136
- 806 42. Marschner, P, Yang, CH, Lieberei, R, Crowley, DE (2001) Soil and  
807 plant specific effects on bacterial community composition in the  
808 rhizosphere. Soil Biol Biochem 33: 1437–1445
- 809 43. McCaig, AE, Glover, LA, Prosser, JI (2001) Numerical analysis of  
810 grassland bacterial community structure under different land  
811 management regimens by using 16S ribosomal DNA sequence  
812 data and denaturing gradient gel electrophoresis banding patterns.  
813 Appl Environ Microbiol 67: 4554–4559
- 814 44. McGonigle, TP, Fitter, AH (1990) Ecological specificity of vesicular  
815 arbuscular mycorrhizal associations. Mycol Res 94: 120–122
- 816 45. Morgan MF (1941) Chemical soil diagnosis by the universal soil  
817 testing system. In: Connecticut Agricultural Experimental Station  
818 Bulletin Number 450
- 819 46. O’Sullivan, AM (1982) The lowland grasslands of Ireland.  
820 In: White J (Ed.) Studies on Irish Vegetation. Royal Dublin  
821 Society, Dublin
- 822 47. Polz, MF, Cavanaugh, CM (1998) Bias in template-to-product  
823 ratios in multitemplate PCR. Appl Environ Microbiol 64:  
824 3724–3730
- 825 48. Ritz, K, McNicol, JW, Nunan, N, Grayston, S, Millard, P,  
826 Atkinson, D, Gollotte, A, Habeshaw, D, Boag, B, Clegg, CD,  
827 Griffiths, BS, Wheatley, RE, Glover, LA, McCaig, AE, Prosser, JI  
828 (2004) Spatial structure in soil chemical and microbiological  
829 properties in an upland grassland. FEMS Microbiol Ecol 49:  
830 191–205
- 831 49. Rodwell, JS (1992) British Plant Communities: Grasslands and  
832 Montane Communities. Cambridge University Press, Cambridge
- 833 50. Rovira, AD (1969) Plant root exudates. Bot Rev 35: 35–37
- 834 51. Schutter, ME, Sandeno, JM, Dick, RP (2001) Seasonal, soil type,  
and alternative management influences on microbial communities 835  
of vegetable cropping systems. Biol Fertil Soils 34: 397–410 836
52. Sparling, GP, Tinker, PB (1978) Mycorrhizal infection in pennine 837  
grassland. I. Levels of infection in the field. J Appl Ecol 15: 943– 838  
950 839
53. Steer, J, Harris, JA (2000) Shifts in the microbial community in 840  
rhizosphere and non-rhizosphere soils during the growth of 841  
*Agrostis stolonifera*. Soil Biol Biochem 32: 869–878 842
54. Thalmann, A (1968) Zur methodik ber Bestimmung der dehydro- 843  
genaseaktivität im Boden mittels triphenyltetrazoliumchlorid 844  
(TTC). Landwirtsch Forsch 21: 249–258 845
55. Thomas, GW (1996) Soil pH and soil acidity. In: Sparks, DL, Page, 846  
AL, Helmke, PA, Loeppert, RH, Soltanpour, PN, Tabatabai, MA, 847  
Johnston, CT, Sumner, ME (Eds.) Methods of Soil Analysis: Part 848  
3, Chemical Methods. Soil Science Society of America, Madison, 849  
WI, pp 475–490 850
56. van Wintzingerode, F, Gobel, UB, Stackebrandt, E (1997) 851  
Determination of microbial diversity in environmental samples: 852  
pitfalls of PCR-based rRNA analysis. FEMS Microbiol Rev 21: 853  
213–229 854
57. Vandenkoornhuysse, P, Ridgway, KP, Watson, IJ, Fitter, AH, 855  
Young, JPW (2003) Co-existing grass species have distinctive 856  
arbuscular mycorrhizal communities. Mol Ecol 12: 3085–3095 857
58. Wardle, DA (1992) A comparative assessment of factors which 858  
influence microbial biomass carbon and nitrogen levels in soil. 859  
Biol Rev 67: 321–358 860
59. Wardle, DA, Bonner, KI, Nicholson, KS (1997) Biodiversity and 861  
plant litter: experimental evidence which does not support the 862  
view that enhanced species richness improves ecosystem function. 863  
Oikos 79: 247–258 864
60. Webster, G, Embley, TM, Prosser, JI (2002) Grassland manage- 865  
ment regimens reduce small-scale heterogeneity and species di- 866  
versity of  $\beta$ -proteobacterial ammonia oxidizer populations. Appl 867  
Environ Microbiol 68: 20–30 868
61. Westover, KM, Kennedy, AC, Kelley, SE (1997) Patterns of 869  
rhizosphere microbial community structure associated with co- 870  
occurring plant species. J Ecol 85: 863–873 871
62. White, TJ, Bruns, T, Lee, S, Taylor, J (1990) Amplification and 872  
direct sequencing of fungal ribosomal RNA genes for phyloge- 873  
netics. In: Innis, MA, Gelfand, DH, Sninsky, JJ, White, TJ (Eds.) 874  
PCR Protocols: A Guide to Methods and Applications. Academic 875  
Press, New York, pp 315–322 876
63. Williams, BL, Grayston, SJ, Reid, EJ (2000) Influence of synthetic 877  
sheep urine on the microbial biomass, activity and community 878  
structure in two pastures in the Scottish uplands. Plant Soil 225: 879  
175–185 880
64. Zhu, YG, Laidlaw, AS, Christie, P, Hammond, MER (2000) The 881  
specificity of arbuscular mycorrhizal fungi in perennial ryegrass– 882  
white clover pasture. Agric Ecosyst Environ 77: 211–218 883

## AUTHOR QUERY

**AUTHOR PLEASE ANSWER QUERY.**

Q1. Figures 1 and 2 are pixelated. Please provide better quality image.

UNCORRECTED PROOF